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OPEN K-variant *BCHE* and pesticide exposure: Gene-environment interactions in a case-control study of Parkinson's disease in Egypt

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Pesticide exposure is associated with increased risk of Parkinson's disease (PD). We investigated in Egypt whether common variants in genes involved in pesticide detoxification or transport might modify the risk of PD evoked by pesticide exposure. We recruited 416 PD patients and 445 controls. Information on environmental factors was collected by questionnaire-based structured interviews. Candidate single-nucleotide polymorphisms (SNPs) in 15 pesticide-related genes were genotyped. We analyzed the influence of environmental factors and SNPs as well as the interaction of pesticide exposure and SNPs on the risk of PD. The risk of PD was reduced by coffee consumption [OR = 0.63, 95% CI: 0.43-0.90, P = 0.013] and increased by pesticide exposure [OR = 7.09, 95% CI: 1.12-44.01, P = 0.036]. The SNP rs1126680 in the butyrylcholinesterase gene BCHE reduced the risk of PD irrespective of pesticide exposure [OR = 0.38, 95% CI: 0.20-0.70, P = 0.002]. The SNP rs1803274, defining K-variant BCHE, interacted significantly with pesticide exposure (P = 0.007) and increased the risk of PD only in pesticide-exposed individuals [OR = 2.49, 95% CI: 1.50-4.19, P = 0.0005]. The K-variant BCHE reduces serum activity of butyrylcholinesterase, a known bioscavenger for pesticides. Individuals with K-variant BCHE appear to have an increased risk for PD when exposed to pesticides.

Parkinson's disease (PD) is a progressive neurodegenerative disease with intraneuronal aggregation of alpha-synuclein, and characteristic motor and non-motor symptoms, affecting more than 6.2 million people globally¹. Monogenic inheritance only accounts for a small proportion of PD cases, whereas the etiology in more than 90% of the patients appears as complex interplay of multiple genetic and environmental risk factors². Knowledge about causative factors is of utmost relevance to develop preventive measures and disease-modifying therapies.

In the 1980s, 1-methyl-4-phenylpyridinium (MPP+) was discovered to induce neurodegeneration and parkinsonism in drug abusers³. MPP+ was marketed as pesticide under the tradename cyperquat⁴ and showed structural similarities to other known pesticides, e.g. paraquat. This finding triggered intensive research about potential

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links between pesticides and PD. Three meta-analyses of epidemiological studies investigating the association of pesticides and PD have been carried out so far^{5,6}. They consistently concluded that pesticide exposure and factors related to pesticide exposure (e.g. rural living, farming or well water drinking) are positively associated with the risk to develop PD^{5,6}.

Recent studies investigated whether risk conferred by pesticide exposure is modified by single nucleotide polymorphisms (SNPs) in candidate genes involved in detoxification or neuronal uptake of pesticides (e.g. aldehyde dehydrogenase 2 (*ALDH2*)⁷, cytochrome P450 2 D6 (*CYP2D6*)⁸, manganese-dependent superoxide dismutase (*MnSOD*)⁹, nitric oxide synthase 1 (*NOS1*)¹⁰, NAD(P)H dehydrogenase [quinone] 1 (*NQO1*)⁹, multidrug resistance protein 1 (*MDR1*)¹¹, glutathione-S-transferase (*GST*)¹², paraoxonase 1 (*PON1*)¹³, dopamine transporter (*SLC6A3*)¹⁴. Some studies found an interaction of pesticide exposure with genetic variants in *ALDH2*⁷, *CYP2D6*⁸, *NOS1*¹⁰, *PON1* concerning PD risk¹³. Most studies, however, did not find significant gene-environment interactions. A recent study searched genome-wide for genetic modifiers of PD risk conferred by pesticides in a relatively small number of patients, without finding any significant results¹⁵.

In Egypt, pesticides are used extensively and under low safety standards, including types of pesticides that have been banned in Western countries for many years due to safety concerns 16 . Presently, around 40% of the Egyptian workforce is employed in agriculture with high likeliness of pesticide exposure 17 . Furthermore, Egypt has an age-adjusted prevalence (\geq 50 years) of 2,500-2,750 PD cases per 100,000 in distinct governorates 18,19 , with a three-fold excess in rural over urban residence, which represents a massive increase by international comparison, particularly also compared to surrounding Arab countries 20 . Therefore, we collected a case-control sample in Egypt to study the association of PD with exposure to pesticides and their interaction with genetic variants involved in pesticide metabolism. Genes and variants of interest were selected by a detailed literature research on genes important for pesticide detoxification with a possible relation to neurodegeneration.

Results

Participants' characteristics. The study sample consisted of n = 416 unrelated PD patients and n = 445 unrelated healthy controls of Egyptian ancestry (Table 1). The sex distribution did not differ between the groups, but PD patients were older than controls. Age at diagnosis, disease duration, and Hoehn and Yahr stage distribution for the PD patients are shown in Table 1.

Environmental factors affect the risk of PD. In a first exploratory comparison, we analyzed differences in single environmental factors between PD patients and control individuals (Table 1). Coffee consumption was the only factor associated with a decreased risk of PD. Factors associated with increased risk of PD were age, well water drinking, illiteracy, use of pesticides at work, and specifically the use of herbicides at home and/or at work. The use of insecticides at home and/or at work showed a trend towards a positive association with PD.

We then constructed a logistic regression model to assess the composite influence of environmental factors on the risk for PD. We considered all factors, which were significantly associated with PD in the single factor analysis (Table 1) for the model.

Well water drinking was significantly associated with PD, but 25 out of 27 well water drinkers (93%) were also exposed to pesticides, leading to high collinearity (P < 0.001) between these factors, with pesticide exposure rather than well water drinking being the likely causal risk factor for PD^{2,5}. Illiteracy was also highly collinear with pesticide exposure (P < 0.001), with 85% of the illiterate participants being pesticide exposed. Again, pesticide exposure rather than illiteracy was the biologically plausible risk factor^{2,5}. Therefore, well water drinking and illiteracy were excluded from the logistic regression model.

In our cohort, 79% of the coffee drinkers were not pesticide-exposed. Thus, coffee consumption was inversely correlated with pesticide exposure (P < 0.001). Nevertheless, we included coffee in the regression model because it is a well-known protective factor for PD^{2,5}. Age was expectedly a highly significant risk factor for PD (P < 0.001). In addition, age interacted with pesticide exposure (P = 0.031) as explanatory variables for PD risk. Age was a somewhat stronger risk factor in the pesticide exposed subgroup (OR for 10 years age difference in the exposed subgroup: 2.95, 95% confidence interval: 2.39–3.71; in the unexposed subgroup: 2.56, 95% confidence interval: 1.18–3.02).

Consequently, the final logistic regression model contained pesticide exposure, coffee consumption, age, and the interaction (age x pesticide exposure). In this analysis, age and pesticide exposure were confirmed as significant risk factors for PD, while coffee consumption was protective (Table 2).

Influence of protective measures on PD risk. To identify factors modulating the risk for PD caused by occupational pesticide exposure, we compared protective measures in the subgroups of occupationally exposed participants, all of whom worked with pesticides in agriculture (n = 156 overall, n = 87 PD, n = 69 controls). Most of them had worked for more than 20 years with pesticides (90.8% of PD patients, 91.3% of controls). Their risk for PD was significantly reduced by wearing gloves during work and by washing hands after work, but not by changing clothes and taking a shower after work (Table 3).

Effect of variants in *BCHE* **on PD risk.** Next, we investigated the influence of genetic factors and geneenvironment interactions on the risk to develop PD. After marker- and sample-wise quality control of the genotyping data, n = 372 PD patients and n = 394 control individuals remained, of whom n = 275 (n = 147 PD, n = 128controls) had been exposed to pesticides. We expanded the logistic regression model described above by the SNP data and by an interaction term (SNP × pesticide exposure). We analyzed the dominant model for all SNPs (Supplementary Table S1).

Characteristic	PD cases (n=416)	Controls (n=445)	OR	95% CI	P-Value
Sex (male) [n (%)]	253 (60.8)	263 (59.1)	1.074	(0.82-1.41)	0.608
Age (y) [mean (SD)]	58.4 (8.8)	48.6 (11.8)	2.56*	(2.18-3.02)	<0.001*
Age at diagnosis (y) [mean (SD)]	55.0 (8.8)	n.a.			
Disease duration (y) [mean (SD)]	3.5 (2.8)	n.a.			
Hoehn & Yahr stage					
Stage 1 [n (%)]	121 (29.1)	n.a.			
Stage 1.5 [n (%)]	111 (26.7)	n.a.			
Stage 2 [n (%)]	107 (25.7)	n.a.			
Stage 2.5 [n (%)]	20 (4.8)	n.a.			
Stage 3 [n (%)]	40 (9.6)	n.a.			
Stage 4 [n (%)]	14 (3.4)	n.a.			
Stage 5 [n (%)]	3 (0.7)	n.a.			
Head trauma [n (%)]	10(2.4)	19 (4.3)	0.55	(0.24-1.17)	0.135
Rural living					
Most of life $[n (\%)]$	142 (34.1)	135 (30.3)	1.19	(0.89-1.59)	0.233
Childhood [n (%)]	138 (33.2)	135 (30.3)	1.14	(0.86-1.52)	0.372
Well water drinking [n (%)]	27 (6.5)	0 (0.0)	n.a.	n.a.	<0.001‡
Illiteracy [n (%)]	129 (31.0)	100 (22.5)	1.55	(1.14-2.11)	0.005
Coffee consumption [n (%)]	66 (15.9)	134 (30.1)	0.44	(0.31-0.61)	< 0.001
Black tea consumption [n (%)]	391 (94.0)	418 (93.9)	1.01	(0.58-1.77)	0.972
Cigarette smoking [n (%)]	69 (16.6)	78 (17.5)	0.94	(0.65-1.33)	0.714
Shisha smoking [n (%)]	39 (9.4)	31 (7.0)	1.38	(0.85-2.27)	0.198
Pesticide exposure#	163 (39.2)	148 (33.2)	1.29	(0.98 1.71)	0.071
Use of pesticides vs. non-exposed	•			•	•
At home only $[n (\%)]$	44 (14.8)	48 (13.9)	1.08	(0.69-1.68)	0.745
At work [n (%)]	87 (25.6)	69 (18.8)	1.48	(1.04-2.12)	0.032
Pesticide classes used at home or/a	nd at work vs. no	on-exposed			
Insecticides only [n (%)]	82 (24.5)	104 (25.9)	0.93	(0.66-1.29)	0.926
Herbicides and other $[n (\%)]$	46 (15.4)	11 (3.6)	4.91	(2.58-9.94)	< 0.001

Table 1. Participants' characteristics. Odds ratio (OR), 95% confidence interval (CI) and *P*-value were calculated using logistic regression; *OR and CI calculated by logistic regression analysis for the ~10-year age difference between cases and controls. "Ever use of pesticides at home or at work, or >50% lifetime residence in rural areas. *Calculated using Fisher's exact test because the logistic regression did not converge. The odds ratio for well water drinking could not be calculated because none of the controls drank well water. n.a., not applicable.

Variable	OR	95% CI	P-Value
Age	2.86+	(2.31-3.61)	2×10^{-16}
Coffee consumption	0.63	(0.43-0.90)	0.013
Pesticide exposure*	7.09	(1.12-44.01)	0.036

Table 2. Environmental factors affecting the risk for PD. Statistical analysis was carried out by logistic regression (see results section for details of the model) using the model formula: affection status \sim pesticide exposure + coffee + age + (age * pesticide exposure); $^+$ OR and CI calculated by logistic regression analysis for the \sim 10-year age difference between cases and controls. *Ever use of pesticides at home or at work, or >50% lifetime residence in rural areas.

Only the SNP rs1126680 in the butyrylcholinesterase gene (*BCHE*) showed significant association with PD *per se* (Table 4; minor allele (G) frequency 0.140 in controls, 0.060 in PD; P = 0.007, OR = 0.38, 95% confidence interval: 0.20–0.70).

SNP rs1803274 was not associated with PD *per se* (minor allele (A) frequency 0.241 in controls, 0.250 in PD), but interacted significantly with pesticide exposure (Table 4; interaction: P = 0.007 dominant). In carriers of the minor allele of rs1803274, pesticide exposure significantly increased the risk of PD (Fig. 1a; P = 0.0005, OR = 2.49, 95% confidence interval: 1.50–4.19) compared to unexposed individuals with the same genotype.

In addition, we performed a subgroup analysis comparing non-exposed participants specifically to participants exposed only to insecticides or to herbicides, respectively. The interaction of rs1803274 was significant in the insecticide only exposed subgroup (Table 4; *P* interaction = 0.002), but not in the herbicide-exposed subgroup (Table 4; *P* interaction = 0.893). In carriers of the minor allele of rs1803274, insecticide exposure increased the

Protective measure	PD cases (n = 87)	Controls (n=69)	OR	95% CI	P-Value		
Wearing gloves during work							
Yes [n (%)]	33 (37.9)	44 (63.8)	0.35	(0.18-0.66)	0.002		
No [n (%)]	54 (62.1)	25 (36.2)					
Washing hands after work							
Yes [n (%)]	67 (77.0)	66 (95.7)	0.15	(0.03-0.47)	0.003		
No [n (%)]	20 (23.0)	3 (4.3)					
Changing clothes after work							
Yes [n (%)]	17 (19.5)	13 (18.8)	1.05	(0.47-2.37)	0.912		
No [n (%)]	70 (80.5)	56 (81.2)					
Shower after work							
Yes [n (%)]	10 (11.5)	15 (21.7)	0.47	(0.19-1.07)	0.088		
No [n (%)]	77 (88.5)	54 (78.3)					

Table 3. Influence of protective measures on the risk for PD. Odds ratio (OR), 95% confidence interval (CI) and *P*-value were calculated using logistic regression.

		SNP main effect			Interaction	
SNP	Model	OR	95% CI	P-Value	P-Value	
Complete sample – pesticide exposed vs. non-exposed						
rs1126680	dominant	0.38	(0.20-0.70)	0.002	0.795	
rs1803274	dominant	0.75	(0.49-1.15)	0.187	0.007	
Insecticide exposed subgroup vs. non-exposed						
rs1803274	dominant	n.a.	n.a.	n.a.	0.002	
Herbicide exposed subgroup vs. non-exposed						
rs1803274	dominant	n.a.	n.a.	n.a.	0.893	

Table 4. Variants in *BCHE* affecting the risk for PD and their interaction with pesticide exposure. Logistic regression analysis of SNP main effect and interaction with pesticides. The interaction *P*-value is derived from the interaction term between SNP and pesticide exposure. The logistic model contained SNP, pesticide exposure, age, coffee consumption and the interaction between pesticides and SNP as well as pesticides and age; affection status \sim SNP + pesticide exposure + coffee + age + (SNP * pesticide exposure) + (age * pesticide exposure); n.a. not applicable for subgroup analysis.

risk of PD (Fig. 1b; P = 0.008, OR = 2.36, 95% confidence interval: 1.26–4.50) in comparison to the unexposed group with the same genotype. Herbicide exposure led to a strongly elevated PD risk independent of genotype (Fig. 1c).

Discussion

Studying 24 SNPs within 15 genes involved in pesticide detoxification and transport, we found one SNP (rs1803274 within *BCHE*) that is associated with increased risk for PD in pesticide-exposed Egyptians. *BCHE* codes for the protein butyrylcholinesterase (BChE) that is alternatively designated as pseudocholinesterase or plasma (choline) esterase. The minor allele of rs1803274 defines the K-variant (Kalow variant) of *BCHE*²¹, which has been shown to reduce the activity of functional BChE in serum by 33%^{22,23}.

Similar to acetylcholinesterase (AChE), BChE hydrolyses choline esters, e.g. the neurotransmitter acetylcholine (ACh). BChE is 10-fold more common in the body than AChE, yet it does not have unique physiological functions that cannot be compensated by other enzymes. It does, however, play an important role as a bioscavenger protecting against organophosphate and carbamate toxicity²⁴. These pesticides prevent degradation of ACh thus causing its accumulation and overstimulation of nerves and muscles with resulting toxic effects²⁵. By binding to pesticides, BChE reduces the amount of active substances that can interfere with AChE to induce acute toxicity, or other esterases (e.g. neuropathy target esterase) to induce chronic neurotoxicity²⁶.

In our study, insecticides but not herbicides significantly increased PD risk in carriers of the K-variant of *BCHE*. Insecticides used in Egypt are mainly organophosphates (e.g. chlorpyriphos) and carbamates (e.g. carbofuran) that are insufficiently "bioscavenged" by the K-variant of *BCHE* presumably explaining the observed increased risk for PD. In contrast, herbicides are mainly pyrimidines (e.g. bispyribac) and organochlorines (e.g. acetochlor) that do not interact with BChE.

One previous study reported an increased number of individuals with homozygosity for K-variant *BCHE* among PD patients compared to age-matched controls $(P=0.051)^{27}$. This finding, however, has not been confirmed so far²⁸. In our sample, K-variant *BCHE* was also not associated with an increased risk for PD by itself, but

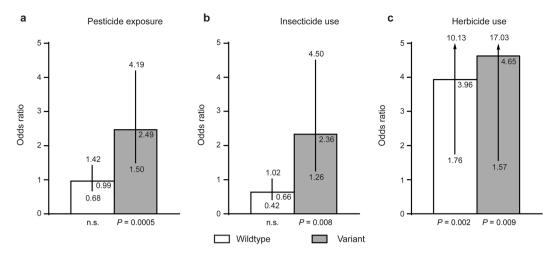


Figure 1. The SNP rs1803274 in *BCHE* increases the risk for PD in pesticide-exposed individuals. Effect of pesticide exposure on PD risk per genotype in the dominant model. The statistical measures are reported with reference to the same genotype (wt = wildtype/wildtype), (var = wildtype/variant or variant/variant) without pesticide exposure for which the odds-ratio is per definition 1. Odds-ratios are found in the upper right corner of the bars and 95% confidence intervals of the odds-ratios are indicated below and above the error bars. (a) Analysis of the whole sample comparing pesticide exposed to unexposed individuals. (b) Analysis of the subgroup of insecticide only exposed subgroup comparing insecticide exposed to unexposed individuals. (c) Analysis of the subgroup of herbicide and other pesticide exposed individuals comparing this group to unexposed individuals.

it possibly facilitated pesticide-induced development of PD owing to the reduced activity and thus less effective bioscavenging property of the K-variant *BCHE*.

Another SNP within *BCHE*, i.e. rs1126680, decreased PD risk in both, pesticide-exposed or unexposed individuals. This is not surprising since rs1126680 does not affect activity and function of BChE even in organophosphate pesticide exposed individuals²⁹. Our findings on BChE, however, are not contradictory. In fact they highlight the different roles and functions of BChE under various conditions. On the one hand, BChE acts as a bioscavenger under pesticide-exposed conditions, backing AChE and protecting the brain against toxic effects²⁶. However, it has recently been discovered that BChE has its own physiological role affecting brain homeostasis³⁰. More important, recent studies proved that BChE might play certain roles in neurodegenerative diseases pathology³¹. However, the functional effects of *BCHE* rs1126680 in this context are unknown so far, but should be elucidated in future investigations.

Since we did not actively match the PD and control groups for age, we assessed the influence of the factors pesticide exposure, coffee consumption, age, and age \times pesticide interaction by logistic regression analysis. This approach confirmed pesticide exposure to increase the risk for PD (Table 2), which is consistent with previous observations in different populations^{2,5}. Our estimate for pesticide exposure (OR = 7.09, 95% confidence interval: 1.12–44.01) is at the upper end of the range reported in prior studies (OR range: 1.5–7.0)⁶ for pesticide exposure as risk factor of PD but the large confidence interval suggests a high degree of uncertainty concerning the exact value. Studying the efficacy of protective measures in participants working with pesticides in agriculture, we found that wearing gloves during work and washing hands after work reduced the risk for PD (Table 3). This is in line with a previous study showing that glove use and hygiene habits are able to reduce the risk of PD associated with certain pesticides³². Additionally, there is convincing evidence that the hands are the most contaminated anatomical region among people working with pesticides³³. Also, it was shown that different pesticides are rapidly absorbed by the skin³⁴ emphasizing that glove use can protect from direct pesticide exposure and thus the risk to develop PD.

Furthermore, we found the well-established protective effect of coffee against PD in the present study as well. In contrast to previous reports, however, we did not find a protective effect of tobacco smoking. This might be due to a possible pesticide contamination of tobacco products in Egypt²⁶.

We also identified a higher rate of illiteracy among PD patients as compared to controls. This is consistent with a previous door-to-door study in an Egyptian governorate that revealed a crude prevalence rate of PD of 1,103/100,000 among illiterates, as opposed to 557/100,000 in the general population 18. Such correlation has not been found in other studies that were mainly conducted in highly industrialized nations. Given that illiteracy was collinear to pesticide exposure in our study, a high degree of illiteracy in pesticide-exposed peasants and less strict adherence to safety measures in this poorly educated group might partially explain the increased risk of illiteracy in Egyptian PD patients. Furthermore, the previously described increased risk for PD in people drinking well water was collinear with pesticide exposure. Therefore, illiteracy and well water drinking are most likely indicators for pesticide exposure in our sample.

The present study confirms pesticide exposure as a risk factor and coffee consumption as a protective factor for PD in an Egyptian population. rs1126680 in *BCHE* decreased the risk for PD regardless of pesticide exposure,

and rs1803274 in *BCHE* (K-variant) increased the risk for PD in individuals exposed to pesticides, particularly to insecticides, such as organophosphates and carbamates. This finding provides a basis to identify persons at risk for individualized preventive measures.

Methods

Ethics approval. The present study was approved by the ethics committee of Mansoura University, Egypt and the Technical University of Munich, Germany and conducted in accordance with the Declaration of Helsinki and all relevant guidelines and regulations. All study subjects provided written informed consent.

Study population. PD patients and controls without neurodegenerative disease were enrolled between January 2013 and December 2015 from the collaborating Neurology Departments of the Universities Mansoura, Ain Shams, Assiut, Sohag, Tanta and Zagazig. Participants underwent a standardized clinical assessment by consultant neurologists specialized in movement disorders. Patients with PD were diagnosed using the UK Brain Bank Criteria³⁵. Patients with atypical, secondary or familial forms of Parkinsonism or other neurodegenerative diseases were excluded. The modified Hoehn & Yahr stage was ascertained in the on-medication state. Controls without neurodegenerative diseases, as ascertained by history and neurological examination, were recruited from attendants of the collaborating hospitals (healthy visitors or patients without neurodegenerative diseases).

Questionnaire data collection. Data about environmental factors assumed to modify the risk of PD was collected by trained study assistants in structured interviews using a standardized questionnaire. The questionnaire included the following sections: General information (sex, age, date of birth, ethnicity), disease history (year of diagnosis, disease duration, medication, family history), residence history (duration of rural or urban living), education (literacy, years of education), occupation history (occupation learnt, working history), nutrition habits (coffee, black tea), smoking habits (years and quantity of smoking, cigarette or shisha use), pesticides used at home or at work (duration, frequency, type of pesticides (insecticides, fungicides, herbicides), and pesticide handling (safety precautions, hygienic measures). Some items of the questionnaire were adapted from the risk factor questionnaires of the National Institute of Neurological Disorders and Stroke (NINDS, www.commondata-elements.ninds.nih.gov/pd.aspx#tab=Data-Standards). Other factors were added to the questionnaire because of their prior epidemiological association with PD^{2,5}. Participants were considered as pesticide-exposed if pesticides were ever used at home or at work or if they resided in a rural area for more than 50% of their lifetime.

Sample preparation and genotyping. Blood-cell-derived genomic DNA (80–100 ng/μL) was genotyped with the EP1 platform on 96.96 Dynamic Array and read by Fluidigm EP1 Genetic Analysis Scanner (Fluidigm Corporation, San Francisco, CA). Twenty-four Candidate SNPs were chosen in genes related to pesticide detoxification [CYP1B1 (rs1056836)³⁶, CYP2B6 (rs3745274)³⁷, CYP2C9 (rs1799853, rs1057910)³², CYP2C18 (rs2296680)³⁸, CYP2E1 (rs2070676)³⁹, PON1 (rs662, rs854560, rs854572)⁴⁰, GSTO1 (rs11191972, rs4925)⁴¹, GSTO2 (rs2297235, rs156697)⁴¹, NAT1 (rs5030839, rs4987076)^{42,43}, NAT2 (rs15561)⁴³, NQO1 (rs1800566)⁹, COMT (rs4680)³⁶, BCHE (rs1803274, rs1799807, rs1126680)^{24,29,44}, PLA2G6 (CM1211192)⁴⁵] and pesticide transport [SLC6A3 (rs27072, rs2550956)^{14,46}]. Assays include tagged, allele-specific PCR forward-primers and a common reverse primer. Genotypes were determined by using allele specific fluorescent probes (FAM and HEX), which were detected by the EP1 scanner. Data was analyzed by the Fluidigm SNP Genotyping Analysis Software to obtain genotype calls. Automatic calls that did not appear clear were either amended manually or uncalled. The overall call confidence was ≥98.5%.

Quality control of genetic data. Genetic data quality control was performed using PLINK 1.9 (www. cog-genomics.org/plink1.9/). First all samples were excluded in which >10% of genotypes were not reliably determined. Subsequently all markers that could not be genotyped in >10% of samples, markers with a minor allele frequency <1% and markers with a Hardy-Weinberg P-value <0.0001 were removed. Quality control of genetic data led to a reduction in sample size from 861 in the analysis of demographic and environmental data to 766 samples included in the analysis of genetic data.

Statistical analysis. Statistical analyses were performed with the statistics software R version 3.3.3 (www.r-project.org). Numbers of study participants in Table 1 were obtained by tabulation. Numbers of study participants positive/negative for more than one variable as found in the results section were calculated by cross-tabulation. Most data were analyzed using logistic regression analyses with models specified in the results section. Logistic regression was performed in base R using the "glm" command, the family "binomial", and the link function "logit". Interaction of variables was analyzed by introducing interaction terms into the logistic regression analysis for the parameters indicated in the results section. *P*-values for interactions are if not stated otherwise the *P*-values of the interaction term. To estimate the *P*-value and odds-ratios for pesticide exposure in the rs1803274 wildtype (wt/wt) and variant carriers (wt/var and var/var) we stratified the sample according to these genotypes and performed a logistic regression for the independent variable pesticide exposure. For demographic and environmental variables, a *P*-value of <0.05 was considered significant. For genetic variables, a *P*-value of <0.01 was regarded as significant. A formal correction for multiple testing was not performed because all analyzed variants were candidate variants already described in the context of pesticide exposure by previous studies.

Data Availability

The data that support the findings of this study are available from the corresponding author.

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Author Contributions

G.U.H., T.W.R., M.S. and G.K. designed the study. T.W.R., M.S., G.K., G.U.H., A.S.S., E.M.K., A.E.-T., G.F., A.E.-M., E.E.-S., M.E.-S., M.E.-G., M.M., M.E.-K., M.A.-N., S.A., U.M., A.D. collected data and carried out data analysis; G.K., A.D., G.U.H., T.W.R. and M.S. conceived and carried out statistical analysis. T.W.R., M.S., G.U.H. and G.K. drafted the manuscript. All authors reviewed and edited the manuscript.

Additional Information

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